**Title:** Trophic phenological mismatch: Disconnects between underlying ecological theory and climate change responses

**Authors:** Heather M. Kharouba1, Elizabeth M. Wolkovich2,3,4

**Affiliations:**

1Department of Biology, University of Ottawa, Canada K1N 9B4

2Organismic and Evolutionary Biology, Harvard University, Cambridge, MA 02138

3Arnold Arboretum of Harvard University, Boston, MA, 02130

4Forest and Conservation Sciences, University of British Columbia, Vancouver, Canada V6T 1Z4

**Email addresses**: heather.kharouba@uottawa.ca; e.wolkovich@ubc.ca

**Correspondence:** Heather M. Kharouba, 30 Marie-Curie, University of Ottawa, Ottawa, ON, Canada, K1N 9B4, 613-562-5800 x6740; heather.kharouba@uottawa.ca

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**Abstract**

Many researchers hypothesize that climate change will lead to phenological mismatches—where the timing of critical life history events between interacting species becomes de-synchronized with potential negative consequences for those interacting species and their ecological communities. Yet, evidence documenting negative impacts on fitness is mixed. The most common ecological theory that underlies these studies is the Cushing match-mismatch hypothesis. It offers a testable hypothesis that—when combined with a system’s pre-climate change baseline condition—can predict the consequences of desynchronized phenology due to climate change. However, a review of the current literature shows that none of the studies collected the data required to provide strong tests of this major hypothesis, making it difficult to assess support for it. Further, 74% of studies fail to define pre-climate change baselines in their study system, making predictions about climate change impacts on consumer fitness difficult. To accurately predict the magnitude and prevalence of mismatches due to climate change, relating empirical observations to underlying mechanisms through hypothesis testing will be required. By adjusting their study designs, researchers can more rigorously test the Cushing hypothesis. We highlight how improved approaches could rapidly advance our mechanistic understanding and thus allow robust predictions of shifts with continuing climate change.

**Introduction**

Climate change is causing phenological shifts—changes in the timing of life history events—that vary across species as well as between functional groups and trophic levels (Ovaskainen et al. 2013; CaraDonna et al. 2014; Thackeray et al. 2016). Such species-specific variation in response to climate change has led to changes in the relative timing of key activities (phenological synchrony) among interacting species (Kharouba et al. 2018). These changes have caused fitness consequences—often termed ‘phenological mismatch’ (Box 1)—and have influenced ecosystem-level properties in some contexts (Post and Forchhammer 2008; Burkle et al. 2013; Plard et al. 2014; Doiron et al. 2015), but not others (Vatka et al. 2011; Burthe et al. 2012). Recent theoretical (Johansson et al. 2015; Bewick et al. 2016) and empirical studies (e.g. Samplonius et al. 2016) based in single systems have worked to improve predictions and address diverse findings. Yet—while there is general agreement that predicting phenological mismatches is critical for determining the extent to which pair-wise species interactions, communities, and ecosystem function (e.g. pollination) will be affected by climate change—we still have no general ability to predict the outcomes of shifts in phenological synchrony due to climate change.

Here, we argue that much of the difficulty in predicting the consequences of climate change-driven shifts in synchrony is due to a disconnect between ecological theory and current empirical approaches used in studies of phenological mismatch. Current methodological inconsistencies across studies and intrinsic differences across systems make it difficult to test the relevant underlying ecological theory in the context of climate change. Without better evidence, we cannot attribute variation in findings of phenological match vs. mismatch across studies to species, site, or more specific mechanisms. Without an understanding of the mechanisms underlying the well-documented patterns in phenological shifts, our ability to make accurate predictions about species’ responses, and species’ interactions, to climate change remains limited (O’Connor et al. 2012; Chmura et al. 2018).

Here we focus on the widely-cited Cushing match-mismatch, or trophic mismatch, hypothesis (1974), the most commonly applied hypothesis concerning consumer-resource interactions in this literature. We show how advances could come from direct tests of the hypothesis and clear definitions of baselines, when possible. Our aim is not to put forward additional hypotheses about the context in which phenological mismatch will occur, which has been reviewed extensively elsewhere (e.g., Miller-Rushing 2010; Renner and Zohner 2018), but rather to help guide the study of phenological mismatch by outlining a path forward to develop robust climate change predictions that can scale up to inference across sites and systems.

Although the Cushing hypothesis has been applied to other types of interactions (e.g. mutualism), we limit our discussion to antagonistic interactions between consumers and their food resources. Below, we provide an overview of the Cushing hypothesis and then outline the divide between the hypothesis and the empirical studies using a systematic literature review of phenological mismatch. We then discuss how current approaches are impeding greater progress in the field, and how changes to research methodologies could rapidly advance our understanding and help forecast of the impacts of climate change on ecological communities—the ultimate goal of most of the phenological mismatch literature.

***Overview of the main ecological theory***

The Cushing match-mismatch hypothesis is the most common ecological theory underlying phenological mismatch studies (Appendix). This hypothesis predicts the often-shown concave-down curve between consumer fitness and relative timing between the consumer and its resource (1974; Figure 1). While this curve has been applied across many ecosystems (e.g. Winder and Schindler 2004; Vatka et al. 2011; Arula et al. 2014), the theory originally emerged from the marine fisheries literature as a way to explain the variation in population recruitment of fish stocks.

Based on life-history theory, the Cushing match-mismatch hypothesis postulates that selective pressure should cause the consumer to temporally ‘match’ the peak of its energetic phase with the peak of resource availability (Figure 1). Given this strong selective pressure, if there is any change to the relative timing of the interaction, there will be a decrease in fitness for the consumer (i.e., a mismatch), thus producing a concave-down curve (Figure 1). For example, if a consumer emerges too early in the spring (Figure 1b) and the resource has not yet emerged, the consumer will risk starvation. At the curve’s limits, consumer fitness should fall to zero when the change in relative timing is sufficiently large (Figure 1). For example, the early-emerging consumer (Figure 1b) can only survive so many days without food.

This hypothesis is based on two important assumptions. First, the resource is the major controller on consumer fitness (i.e., strong interaction strength, bottom-up control; Cury et al, 2003). Second, the consumer and resource must present a certain degree of seasonality (e.g., resource is only available for part of the year; Durant et al. 2005; Durant et al. 2007), limiting the optimal period for growth and reproduction for the consumer. If neither of these assumptions is met, then fitness consequences due to changes in the relative timing of the interaction will be either weak or non-existent.

***Disconnect between theory and empirical studies***

In its original state, the hypothesis has been debated, contested and criticized, particularly in the marine literature (Leggett and DeBlois 1994; Durant et al. 2007). In part because, although a relatively simple hypothesis, it is inherently difficult to test in the field, an assertion even Cushing himself made. Indeed, the shape and strength of the relationship of the curve varies greatly across observational studies (e.g., Philippart et al. 2013; Reed et al. 2013; Plard et al. 2014; Atkinson et al. 2015). While some have suggested that this is because of data limitations and the model’s implication of complex multitrophic dynamics (Durant et al. 2007; Kerby et al. 2012), we argue that there are two key theoretical areas that make it difficult to determine whether this hypothesis is widely supported in the context of climate change: studies often fail to rigorously test the Cushing hypothesis and/or test pre-climate change conditions (Figure 2). Below, we introduce the current objectives of the phenological mismatch literature and then discuss these two areas.

To put our argument in context, we systematically reviewed the phenological mismatch literature. From this search, we examined 43 observational studies that met our criteria (see Appendix for details) for evaluating the effects of a change in relative timing on the fitness or performance (hereafter referred to as performance) of the consumer. The majority of the studies (27/43) focused on: i) documenting how climate change is affecting the timing of a trophic interaction and how those changes have affected the consumer’s performance; and ii) evaluating the roles of abiotic and biotic factors in influencing synchrony. The ultimate goal of these studies was to predict the impacts of climate change on ecological communities (hereafter called ‘climate change’ studies). Our database also included studies (16/43) whose aim was to understand the underlying processes related to timing (e.g., coevolution, life history trade offs, food web dynamics) that drive consumer or resource dynamics (hereafter called ‘fundamental studies’), independently of climate change.

*i) Testing the Cushing hypothesis*

The Cushing hypothesis offers testable predictions of the magnitude and direction of demographic changes in response to climate-change driven shifts in synchrony. To date, much research in the biological impacts of climate change literature has focused on the direct relationships between organisms and the abiotic environment (e.g., Menzel et al. 2006, Chen et al. 2011) rather than testing predictions about the outcomes of species’ responses to changes in abiotic factors (Lavergne et al. 2010; O’Connor et al. 2012; Mouquet et al. 2015; Barner et al. 2018). However, progress on the Cushing hypothesis requires rigorously testing the assumptions and ultimate mechanisms from a diversity of ecological and evolutionary theory (Figure 2). This represents the major challenge of the hypothesis and—we argue—may be why support for it has been so mixed.

Many mechanisms can produce the Cushing curve, and they vary across systems, space and time. Here we focus on the ultimate mechanisms related to life history theory and food web theory. Mechanisms predicted by life-history theory (Box 1), for example, trade-offs between fecundity and mortality, and bet-hedging strategies (Figure 2), can produce the Cushing curve—as consumers maximize their fitness through ideal timing with their primary resource—this forms the theoretical basis for the original hypothesis. The Cushing hypothesis, however, is also connected to food web theory which focuses on predator prey dynamics (Box 1) and is related to the timing of a consumer to its food resource.

Identifying the ultimate mechanisms (e.g., a specific life history trade-off) relates to the likelihood that the key assumptions of the Cushing hypothesis are met (Figure 2). For example, if mechanisms such as density dependence or top-down population regulation are thought to be at play, then the first assumption of the hypothesis is unlikely to be met. Even when both of the assumptions of the Cushing hypothesis have been met, identifying the ultimate mechanism is critical to predicting how interactions will respond to climate change (discussed in more detail in the next section; Figure 2). Given the diversity of theory related to the Cushing hypothesis and therefore potential diversity in the outcomes in consumer performance (e.g., magnitude, direction) due to phenological shifts in synchrony, the better our mechanistic understanding is of these patterns, the more robust our predictions will be (e.g., steepness of the Cushing curve).

Studies deviate in what forces they hypothesize control the peak in the food resource (a critical component of the Cushing curve and a key assumption, see Figure 1 and 2). For example, in aquatic systems—where top-down forces are generally more influential compared to terrestrial systems (Shurin et al. 2005)—many studies suggest that the resource peak is controlled by release from, or predation by, a consumer (Carpenter and Kitchell 1996; Shurin and Seabloom 2005; Borer et al. 2006). If the resource peak is controlled by release from a predator rather than abiotic factors, then the first assumption of the hypothesis is unlikely to be met and the subsequent predictions from the Cushing hypothesis are unlikely to hold. This is a very different hypothesis from others that suggest seasonality in the environment produces the resource peak (Hampton et al. 2006). In terrestrial systems, the relative contribution of direct vs. indirect effects on consumer dynamics remains poorly understood (e.g. Boggs and Inouye 2012) and these two hypotheses can result in different types of effects on population dynamics.

The diversity of ecological theory that can produce the basic shape of the Cushing curve means that different researchers may collect very different data depending on what mechanism they suspect underlies the curve in their particular system or what their main objective is. Researchers studying the hypothesis from the lens of life history theory require per capita estimates of fitness, including measurements of fecundity and mortality, and how they vary across ontogeny; such measurements allow researchers to assess fitness consequences at the level of selection (i.e., the individual) and to directly link any fitness changes to timing. Studies addressing the Cushing hypothesis from the perspective of food web theory require equivalent data on the consumer and resource but do not require per capita fitness consequences for the consumer.

Strong tests of the Cushing hypothesis include clear tests of the assumptions alongside tests of multiple potential mechanisms underlying relationships between phenological match and consumer performance. To accomplish this, equivalent data would ideally be assessed for both the consumer and resource—at the level of the individual when possible. Equivalent data for the consumer and resource facilitate testing assumptions and multiple mechanisms related to the Cushing hypothesis at once. Without at least population-level data on the resource, it is difficult to test the second assumption of the hypothesis; for example, resource abundance can be used to determine the likelihood that consumers will encounter enough resource at the tails of their temporal distributions, thus defining the degree of seasonality in the resource. Per-capita data on the resource, when available, also extend our mechanistic understanding beyond the consumer, thus improving our ability to make predictions of climate change-driven phenological mismatches for the interaction in its entirety. For example, fine-scale data on the resource are needed to fully incorporate feedbacks between the consumer and resource (e.g., Thackeray 2012), measure interaction strength (which informs a key assumption of the hypothesis, Miller-Rushing et al. 2010), and accurately predict changes in resource abundance due to climate change. Together, these data help to determine the wider implications of phenological mismatches for the community (e.g., Nakazawa and Doi 2012; Revilla et al. 2014).

Currently such data are non-existent in this literature (i.e. none of the pair-wise interactions had individual-level data for both the consumer and resource). Of the interactions we examined with per-capita performance data on the consumer (i.e. life history; n=28 interactions), none had equivalent measures of performance on the resource (Table 1). For the majority of these interactions (18/28), the researchers measured performance at the community level for the resource (e.g. biomass; Table 1) and many had no measure at all for the resource (6/28 interactions; Table 1). As expected, researchers that measured population- or community-level data on the consumer (i.e. food-web; n=18 interactions) collected equivalent data on the resource and consumer (e.g. abundance). For the majority of these interactions, researchers also collected population-level (7/18) or community level data for the resource (7/18) (Table 1). Across all life history and food-web interactions, researchers measured per-capita performance on the resource for only one interaction (Philippart et al. 2003).

Part of these discrepancies may stem from differing generation times—studies that can easily observe a resource curve shaped by predation often focus on organisms with generation times on the scale of days to week. For example, aquatic studies that focus on phytoplankton as the resource may observe many generations in one summer while a terrestrial study focused on caterpillars in temperate areas would generally observe a comparably smaller number generations (e.g. 1-2, depending on latitude and species for caterpillars). When populations turn over too quickly to track individuals, researchers struggle to collect robust data on per capita fitness. The fundamental difference in the scale of generation times pervades aquatic/terrestrial comparisons, particularly at lower trophic levels (Borer et al. 2005; Shurin et al. 2005; Gruner et al. 2008), and may drive the difference in the objectives of aquatic vs. terrestrial studies in the phenological mismatch literature. Indeed, the majority of researchers studying aquatic-based interactions in our review approached the Cushing hypothesis from the perspective of food-web theory (15/20 interactions), whereas researchers approached terrestrial-based interactions from life-history theory (23/26; Table 2).

A further divide across studies may come from the size, and population size, of the consumer, properties correlated with generation time: it can be difficult to collect per capita fitness data when consumers are too small and numerous, ultimately making it difficult to track the fate of individuals. In cases where individuals can be distinguished, marked and monitored, it is possible to gather per-capita data; otherwise data are more likely to be aggregated at the population or community-level. Of the life history interactions, the majority of consumers were birds (19/28) and mammals (6/25; Table 2). Comparatively, invertebrates were measured at the population or community level and the vast majority were aquatic (11/12) (Table 2). Therefore, only certain classes of organisms are tested for each category of mechanisms and in one type of biome, further limiting generalizations across systems.

This places the phenological mismatch studies on a continuum: at one end, aquatic studies focus generally on food web mechanisms, collect equivalent data on the consumer and resource but no per-capita fitness data and at the other end, terrestrial studies focus on life history theory from the lens of the consumer, collect individual-level data on the timing and fitness of the consumer, with much less information on the resource. The fundamental problem with these approaches is that researchers test only one piece of the much larger field of mechanisms that could underlie the Cushing curve. Though it is highly possible that both food web and life history theory together explain many of the consumer-resource systems studied, data limitations make it hard to assess mechanisms related to both theories at once. Consequently, most studies do not actually provide strong tests of the Cushing hypothesis, making it difficult to refute the hypothesis if no evidence is found, a prevalent issue in ecology and evolution (Betini et al. 2017). If researchers find their proposed mechanisms explain little of the variation they observe, they rarely have the data to test alternative hypotheses.

Without strong support for the Cushing hypothesis (i.e., high explained variation and clear patterns that align with predictions), the mechanisms underlying it will be uncertain. This uncertainty can be caused by two reasons: (1) hypotheses that are not mutually exclusive (i.e. different mechanisms have the same predictions). For example, the relative timing of an interaction will change in the same direction regardless of whether temperature (or some other abiotic factor) directly or indirectly affects a resource’s peak; or (2) considering multiple mechanisms yields different predictions than considering only a single mechanism. For example, researchers have pointed out how a broader perspective on life history, such as including other sources of mortality beyond those related to resource access or alternative resource types (Samplonius et al. 2016), can lead to alternative predictions than the synchrony predicted by the Cushing curve (Singer and Parmesan 2010; Johannsson & Jonzen 2012).

*ii) Testing pre-climate change conditions*

Applying the Cushing hypothesis to understanding—and ideally predicting—the consequences of climate change for interacting species requires more than testing the assumptions and underlying mechanisms of the hypothesis (Figure 2). If the assumptions of the Cushing hypothesis are met and the mechanisms understood, researchers still need an understanding of the system dynamics before climate change, and the phenological cues that affect the timing of consumer and resource levels, to predict what shifts in the climate have—and will—do to the timing and fitness of the players (Figure 2). Below we discuss these steps in more detail.

*(a) Identifying pre-climate change baselines*

The concept of pre-climate change baseline (i.e., defining the range of natural variation) is rarely, if ever, considered by phenological mismatch studies; yet, establishing a baseline in a system has important implications for the Cushing hypothesis in both climate change and fundamental studies. Of the studies we considered, only 26% (11/43) of the studies began before the early 1980s, the same time that pronounced recent climate change began (Solomon et al. 2007). Even in this subset of studies, not all studies had more than three years of data (2/11 had only one or two years of data for inference). This is also an issue with the phenological synchrony literature (Box 1), which does not necessarily consider the ecological consequences of shifts in synchrony (Kharouba et al. 2018). Even those studies with time-series that extend into decades before the 1980s do not explicitly define a pre-climate change baseline in synchrony (but see Philippart et al. 2003; Adrian et al. 2006; Arula et al. 2014).

There are several major implications of not establishing a pre-climate change baseline (Figure 3). First, studies without a baseline are implicitly assuming conditions of stationarity (i.e., a constant underlying probability distribution) before climate change occurred (Figure 3a; Wolkovich et al. 2014). This is problematic when climate change has led to non-stationarity to different extents in different systems (Figure 3a). Without a defined pre-climate change baseline in the system, we do not know whether, and to what magnitude, we should predict a divergence in synchrony due to climate change (Figure 3a). It is then further difficult to determine how much of the observed change in the interaction can be attributed to climate change, some other driver, for example nitrogen deposition (Edmondson 1994), land use change (Sala et al. 2000), species introductions (e.g., Ricciardi et al. 1998; Fritts and Rodda 1998; Verschuren et al. 2002;), or natural variation. Thus, knowledge of the pre-climate change baseline may rule out climate change and identify other drivers, which is key for the prioritization of conservation management decisions for threatened species.

Second, these studies confound a pre-climate change baseline with the peak of the Cushing curve: that is, they assume that fitness was highest before climate change because the consumer’s timing was well matched to the resource’s (i.e. the synchrony hypothesis; Box 1; Figure 3b; Singer and Parmesan 2010). The resulting prediction is that climate change will necessarily change the relative timing of the interaction (i.e. non-stationarity) leading to a decline in the consumer’s fitness (Figure 2, Figure 3b). However, without knowledge of the pre-climate change baseline, it is difficult to predict the direction and magnitude of fitness consequences due to climate change-driven shifts in synchrony, especially when alternative conditions could be the baseline (Figure 3b,c).

An alternative hypothesis put forward by Singer and Parmesan (2010) suggests that in some contexts, ’adaptive asynchrony’ (i.e. when the most energetically demanding phase of the consumer is not lined up with the peak resource availability for many individuals in the population; Box 1; Figure 3b). They postulate that in some systems, asynchrony in many individuals in a population might be adaptive due to life-history tradeoffs. This could happen when a consumer cannot measure the timing of its resource well (Singer and Parmesan 2010), for example; in this scenario, some individuals would appear mismatched, while others would appear matched. This scenario may persist most often when fitness consequences of mismatch are not extremely high; otherwise it could lead to population extirpation (Figure 3b).

In either hypothesis (i.e. synchrony vs. adaptive asynchrony; Figure 3b), a pre-climate change baseline of asynchrony (Box 1; Figure 3c) could occur through several mechanisms. Shifts in the drivers of phenology for the consumer or resource (e.g., nutrient enrichment) could have pushed the system away from synchrony before climate change. An asynchronous baseline would also be predicted—at times—by a co-evolutionary arms race: when the resource is ahead in the arms-race, asynchrony would the predicted baseline (Figure 3c), this would then shift back towards synchrony when the consumer is winning the arms race. In systems where asynchrony or other transient conditions might be the baseline state, our ability to anticipate the effects of climate change will be even more challenging as these transient conditions could lead to vastly different predictions under climate change (Figure 3c). For example, interactions might become less (i.e. fitness increases) or even more mismatched (i.e. fitness decreases) with climate change (Figure 3b-c).

The pre-climate change baseline is also dependent on the ultimate mechanism(s) generating the Cushing curve (Figure 2). For example, a species that provisions its offspring with resources acquired during the breeding period (i.e., an income breeder; e.g., West Greenland caribou: *Rangifer tarandus*) is predicted to have a pre-climate change baseline of synchrony, whereas a species that provisions its offspring with resources gained prior to reproduction (i.e., a capital breeder; e.g., muskoxen*: Ovibos moschatus*) is more likely to have a baseline of asynchrony (Kerby and Post 2013). Therefore, researchers need an understanding of the system dynamics before climate change began.

Establishing a pre-climate change baseline also has implications for fundamental studies that aim to understand the underlying processes of consumer-resource dynamics. A pre-climate change baseline would determine whether the system is most probably in equilibrium or in transient dynamics. For example, life-history trade-offs, including those that can cause patterns predicted by the Cushing hypothesis, predict some form of equilibrium conditions, but if climate change has pushed the system away from a previous baseline, then the system might likely be in a transient phase rather than at equilibrium (Figure 3). To this end, an understanding of the system’s pre-climate change baseline state is important for providing context and even designing fundamental studies.

*(b) Identifying proximate cues*

For climate change to lead to phenological mismatch, the underlying assumption is that changes in climate will drive changes in the relative timing of species interactions. Therefore, predicting how climate change will affect the relative timing of a consumer and resource, and potentially the fitness of the consumer, will also depend on the identification of the proximate phenological cues of the consumer and resource (Figure 2). For example, climate change is likely to lead to fitness declines for a consumer that has a pre-climate change baseline of synchrony and that shares different environmental cues with its resource (Figure 2). Currently, relatively little is known about the similarity of cues for interacting species across trophic levels (Chmura et al. 2018). Finally, predictions will also depend on how phenological cues will change under climate change scenarios (Chmura et al. 2018).

**Towards robust forecasting of phenological mismatch**

While the Cushing hypothesis provides testable predictions regarding the consequences of phenological mismatch in response to climate change, it is uncertain how much support there is for it. Here we have suggested two major reasons for this, both of which provide a clear path forward for progress in the field (Figure 2). First, most studies do not actually provide strong tests of the Cushing hypothesis because (a) studies do not collect equivalent fitness and population data on the consumer and resource, and (b) multiple mechanisms are not evaluated. This makes it difficult to refute the Cushing hypothesis if no evidence is found. Unless researchers are extremely clear about the mechanistic hypotheses they are testing, progress will continue to be slow. Second, pre-climate change baselines are not defined in the majority of studies in this literature. Lack of a relevant baseline makes it difficult to determine whether recent phenological mismatch can be attributed to climate change and limits our ability to understand and predict the direction and magnitude of phenological mismatch due to climate change.

Moving the field of phenological mismatch forward (i.e., from documentation towards forecasting) requires higher quality data that enables the testing of fundamental hypotheses and defining of key baselines. Below, we highlight how new approaches can improve testing of the Cushing hypothesis. These approaches will also provide the framework for building the required depth in evidence across studies to determine general quantitative patterns in phenological mismatch and their underlying mechanisms, an approach not currently possible given the state of the field.

*(i) Testing fundamental theory*

By having the potential to test multiple mechanisms and define key baselines, the integration of long-term data (i.e. that start before the1980s) and an experimental approach in a single system represents the ‘gold standard’ to move the field of phenological mismatch forward. However, in the absence of long-term data, the integration of observational data across a spatial gradient in climatic conditions, and thus variation in the relative timing of the interaction, with an experimental study can provide a powerful approach to test multiple hypotheses. For example, Samplonius et al. (2016) combine descriptive and experimental data to test many confounding factors and key assumptions of the Cushing hypothesis in their system (e.g., ontogenetic dietary shifts, habitat quality). Consequently, they advanced the field by providing a strong test of the Cushing hypothesis, demonstrating that this hypothesis is pertinent in their system, and ruling out competing hypotheses related to habitat quality.

Independent of whether observational data is available, experiments can be used to test and disentangle multiple mechanisms, for example direct from indirect effects or abiotic mismatches from phenological ones(e.g., Bauerfeind and Fischer 2013; Rudolf and Singh 2013). Experiments provide a high degree of inference about mechanism that descriptive data alone cannot provide. For example, Berger et al. (2014) manipulated water temperature, stratification depth and the presence/absence of zooplankton in lake mesocosms to separate the direct effects of light and temperature on spring plankton dynamics from effects mediated through the consumer. They showed that temperature had little effect on zooplankton phenology and performance suggesting that climate change is unlikely to result in phenological mismatch in this system. The extrapolation of their results to the lake ecosystem are uncertain, however, given the lack of long-term data in the system and the uncertainty about phenological cues for *Daphnia* in the field (Berger et al. 2014).

Baselines themselves can also be used to test multiple mechanisms. Researchers with long-term data can define specific hypotheses about the expected vs. observed patterns of environmental change in their systems and then attribute variation in phenological mismatch to the main driver. For example, many lakes experienced large changes in nutrient inputs in the 1960s and 1970s (e.g., George 2012; Lake Washington (Edmondson 1994)), before the onset of recent climate change in the early 1980s. Researchers can test different breakpoints with their long-term data to determine where the non-linearity most likely began, thus providing evidence for which factor is likely to be the major driver of changes in phenological synchrony. For example, studies have shown that changes in nutrient inputs have had larger impacts on plankton dynamics in Lake Washington than warming over the past three decades (Law et al. 2009; Francis et al. 2014).

Higher temporal resolution data in some systems may help with teasing apart different mechanisms and testing the assumptions of the Cushing hypothesis. Regardless of aim, ideally, sampling frequency should be relative to the life history of the species of interest. For example, testing any mechanism related to life history theory requires data about ontogeny. In the great tit-winter moth system, caterpillar biomass sampling is usually conducted a few times a week (e.g., Vatka et al. 2014). Given that the larval stage of the winter moth life cycle is typically 4-6 weeks (Holliday 1977; Tikkanen et al. 2000), this sampling frequency allows researchers to obtain an accurate estimate of growth and development for this life cycle stage. In the lower trophic levels of aquatic systems, where there is often quick turnaround between producers and consumers, it can be difficult to determine how strong of an influence producers have on consumers and vice versa (Carpenter and Kitchell 1996). Without higher temporal resolution data, it is difficult to assess whether this key assumption of the Cushing hypothesis (i.e., the dependence of the consumer on the producer) is supported. In cases where there are multiple short, potentially overlapping, generations, determining the seasonal order of the consumer vs. producer is necessary to assess the degree of dependence of the consumer on the producer.

Higher temporal resolution data will also help determine the seasonal availability of the consumer and resource, the other key assumption of the Cushing hypothesis (Figure 1b-d). For example, temperate aquatic and terrestrial systems differ in how many organisms survive the winter. While many terrestrial plants and herbivores have dormancy phases during the winter, in aquatic systems some zooplankton are never in complete dormancy (i.e., they remain at low densities), and thus do not fully meet the assumption of the hypothesis. Without more frequent data, accurately estimating the seasonal distribution of the lower trophic levels of some aquatic systems will be difficult. Nevertheless, some aquatic studies have been successful in achieving daily sampling frequency of lower trophic levels (e.g., Wiltshire et al. 2008), suggesting that—though more challenging—critical assumptions of the Cushing hypothesis can be tested in aquatic systems.

*ii) Defining baselines*

Researchers with long-term data have the potential to test whether synchrony vs. asynchrony was the pre-climate change baseline (Figure 2; Figure 3). By testing for a clear peak in fitness across years in relation to the relative timing of the interaction (i.e. evidence for the synchrony hypothesis) in the pre-climate change data, researchers can place their system on the Cushing curve (Figure 3). Whenever possible, incorporating the most energetically expensive part of the consumer’s phenology or estimating when the measured phenological phase occurs relative to the most expensive part (e.g., food demands of the great tit (*Parus major*)) chicks are highest 9 days after hatching, the phenological phase most commonly monitored; Gebhardt-Henrich 1990; Keller and van Noordwijk 1994; Visser et al. 2006) will help define a peak in consumer fitness. If there is no clear peak in fitness, then well-designed experiments (in systems where possible) that have the ability to reveal the limits of the curve (i.e. where consumer fitness reaches zero) can be used to reject the hypothesis of an asynchrony baseline. Knowledge of when/if different resources are available (e.g. Samplonius et al. 2016) and whether the interaction type changes throughout the season (add example from meta-analysis) can also help describe the limits of the curve.

Where long-term data are not available, there is no perfect solution to the challenge of defining a baseline. Researchers with multiple years of data may be able to use inter-annual variation in abiotic conditions (e.g., temperature) to define a baseline, where closer-to mean conditions represent a proxy for a historical baseline and, for example, cooler vs. warmer years act as alternatives to pre- and post- climate change, respectively (e.g., Gullett et al. 2013; Francis et al. 2014). However, extrapolation from short-term datasets can only work if the species responses’ to varying climate themselves are not in flux, a criterion difficult to meet in many systems already (Visser et al. 2006; Singer and Parmesan 2010). Null modeling to simulate historical data may be another alternative (e.g., Dornelas et al. 2014; Sgardeli et al. 2016; Kharouba et al. 2018). This null model could be used to put the magnitude of observed changes in phenology and performance since significant climate change began into context and to quantify how much variation may be due to noise (i.e., factors other than climate change; Kharouba et al. 2018).

When not all required data are available across a long time period, the integration of experimental and observational data for a single system could again be a useful approach (Figure 4). For example, in one study the performance data for an insect herbivore (winter moth; *Operophtera brumata*), and its main host plant (common oak; *Quercus robur*) are not available for more than 10 years from a single location (van Asch and Visser 2007); however, different researchers manipulated the timing of larval emergence relative to bud break to test for support for the Cushing hypothesis (Tikkanen and Julkunen-Tiitto 2003; Figure 4). Although the strength of the relationship may vary across populations (e.g., Visser et al. 1998; Charmantier et al. 2008), a synthesis of the findings from the two approaches (i.e. observational and experimental) can shed light on the state of our knowledge of the system and can guide future research. Here, for this system, we learned where the interaction is currently placed along the curve, how little of the curve is sampled with a 10 year post-climate change time-series, and that—without a clear baseline—it is difficult to predict how the performance of *O. brumata* will be affected by changes in phenological synchrony. Given the knowledge about the Cushing curve for this interaction, modelling pre-climate change baselines based on knowledge of phenological cues using a hindcasting approach (Senner et al. 2016; Deacy et al. 2017), process-based phenological models (Chuine and Régnière 2017), and/or phenologically explicit consumer-resource models (Bewick et al. 2016) could help to fill in data about the interaction before climate change began. Experiments could help define the parameters of a process-based model which could then elucidate that mechanisms may—or may not—appear feasible for the interaction (Chuine and Régnière 2017).

*Final thoughts on forecasting*

Given the complexity involved, accurately forecasting phenological mismatch in response to climate change is a major test of ecological theory and methods, and requires a mechanistic understanding of the processes involved to ensure robust predictions of changes in consumer performance (e.g., steepness of the Cushing curve). Previous work has identified factors that will help predict *when* a mismatch is more likely to lead to a population decline but this work is largely theoretical. While we have outlined here how to work around data limitations, the best tests of mismatch theory will come from data-rich systems where the energetic links between consumer and resource species are well understood. With such information in hand, researchers can then begin to consider forecasting longer-term demographic responses due to mismatch with continued climate change. Forecasting of phenological synchrony can be built off climate projection forecasts whenever phenological cues for both species (Chuine and Régnière 2017; Chmura et al. 2018) are well known. Forecasting longer-term demographic responses will require data about lifetime fitness for the consumer and comprehensive data on the resource, a standard that, to our knowledge, has yet to be met by any empirical study to date.

As more information is gathered across habitats and diverse consumer-resource interactions, forecasting should eventually move beyond a system-specific approach and towards forecasting diverse systems through a multi-level modeling approach based on our mechanistic understanding of what drives mismatch. Such a model would be built on mechanistic studies of the mismatch hypothesis that carefully measure the fitness, phenologies, interactions and other still-unknown critical attributes of systems. The model would then allow researchers to measure currently unknown but critical attributes of species, sites, and ecosystems that control the shape, strength and uncertainty surrounding the Cushing curve. But progress towards such an approach requires the ability to combine knowledge from across diverse systems into one complex but well-defined mechanistic model. As we have outlined above, most current data in the field make it hard or impossible to compare across systems. Given intrinsic differences between aquatic and terrestrial systems, progress may be accelerated if these systems are considered separately before being combined. Nevertheless, without improved methods – to test multiple mechanisms and define pre-climate change baselines – the goal ofgeneral predictions of the ecological consequences of shifts in phenological synchrony will remain well out of reach.

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**References**

Adrian, R., Wilhelm, S. and Gerten, D.*Life-history traits of lake plankton species may govern their phenological response to climate warming*. Global change biology, **2006**, Vol. 12(4), pp. 652-661

Arula, T., Gröger, J., Ojaveer, H. and Simm, M. *Shifts in the spring herring (Clupea harengus membras) larvae and related environment in the Eastern Baltic Sea over the past 50 years* PloS one, **2014**, Vol. 9(3), pp. e91304.

van Asch, M. and Visser, M. E. *Phenology of forest caterpillars and their host trees: the importance of synchrony* Annu. Rev. Entomol., Annual Reviews, **2007**, Vol. 52, pp. 37-55

Atkinson, A., Harmer, R. A., Widdicombe, C. E., McEvoy, A. J., Smyth, T. J., Cummings, D. G., Somerfield, P. J., Maud, J. L. and McConville, K. *Questioning the role of phenology shifts and trophic mismatching in a planktonic food web*. Progress in Oceanography, **2015**, Vol. 137, pp. 498-512

Barner, A. K., Chan, F., Hettinger, A., Hacker, S. D., Marshall, K. and Menge, B. A.*Generality in multispecies responses to ocean acidification revealed through multiple hypothesis testing*. Global change biology, **2018**, Vol. 24(10), pp. 4464-4477.

Bauerfeind, S. S. and Fischer, K. *Increased temperature reduces herbivore host-plant quality*. Global Change Biology, **2013**, Vol. 19(11), pp. 3272-3282

Berger, S. A., Diehl, S., Stibor, H., Sebastian, P. and Scherz, A. *Separating effects of climatic drivers and biotic feedbacks on seasonal plankton dynamics: no sign of trophic mismatch*. Freshwater Biology,  **2014**, Vol. 59(10), pp. 2204-2220.

Betini, G. S., Avgar, T. and Fryxell, J. M. *Why are we not evaluating multiple competing hypotheses in ecology and evolution?* Royal Society open science, The Royal Society Publishing, **2017**, Vol. 4(1), pp. 160756

Bewick, S., Cantrell, R. S., Cosner, C. and Fagan, W. F. *How resource phenology affects consumer population dynamics.* The American Naturalist, University of Chicago Press Chicago, IL, **2016**, Vol. 187(2), pp. 151-166

Boggs, C. L. and Inouye, D. W. *A single climate driver has direct and indirect effects on insect population dynamics.* Ecology Letters,  **2012**, Vol. 15(5), pp. 502-508

Borer, E., Seabloom, E., Shurin, J., Anderson, K., Blanchette, C., Broitman, B., Cooper, S. and Halpern, B. *What determines the strength of a trophic cascade?* Ecology,  **2005**, Vol. 86(2), pp. 528-537

Borer, E. T., Halpern, B. S. and Seabloom, E. W. *Asymmetry in community regulation: effects of predators and productivity*. Ecology,  **2006**, Vol. 87(11), pp. 2813-2820

Burkle, L. A., Marlin, J. C. and Knight, T. M. *Plant-pollinator interactions over 120 years: loss of species, co-occurrence, and function*. Science, American Association for the Advancement of Science, **2013**, Vol. 339(6127), pp. 1611-1615.

Burthe, S., Daunt, F., Butler, A., Elston, D. A., Frederiksen, M., Johns, D., Newell, M., Thackeray, S. J. and Wanless, S. *Phenological trends and trophic mismatch across multiple levels of a North Sea pelagic food web.* Marine Ecology Progress Series, **2012**, Vol. 454, pp. 119-133.

CaraDonna, P. J., Iler, A. M. and Inouye, D. W. *Shifts in flowering phenology reshape a subalpine plant community*. Proceedings of the National Academy of Sciences,  **2014**, Vol. 111(13), pp. 4916-4921.

Carpenter, S. R. and Kitchell, J. F. *The trophic cascade in lakes*. Cambridge University Press, **1996**

Charmantier A, et al. *Adaptive Phenotypic Plasticity in Response to Climate Change in a Wild Bird Population*. Science*,***2008**, 320:800-803.

Chen, I.-C., Hill, J. K., Ohlemüller, R., Roy, D. B. and Thomas, C. D. *Rapid range shifts of species associated with high levels of climate warming*. Science, American Association for the Advancement of Science, **2011**, Vol. 333(6045), pp. 1024-1026.

Chmura, H. E., Kharouba, H. M., Ashander, J., Ehlman, S. M., Rivest, E. B. and Yang, L. H. *The mechanisms of phenology: the patterns and processes of phenological shifts*. Ecological Monographs,  **2018**

Chuine, I. and Régnière, J. *Process-based models of phenology for plants and animals*/ Annual Review of Ecology, Evolution, and Systematics, Annual Reviews, **2017**, Vol. 48, pp. 159-182

Cury, P, Shannon, L and Shin, YJ. *The functioning of marineecosystems: a fisheries perspective.* Responsible fisheries in the marineecosystem. FAO, Rome and CABI Publishing, **2003**, pp. 103–123

Cushing, D. H. *The natural regulation of fish populations*. HardenJones, F. R. *(ed.)* Sea Fisheries Research. Elek Science, **1974**, pp. 399-412

Deacy, W. W., Armstrong, J. B., Leacock, W. B., Robbins, C. T., Gustine, D. D., Ward, E. J., Erlenbach, J. A. and Stanford, J. A. *Phenological synchronization disrupts trophic interactions between Kodiak brown bears and salmon*. Proceedings of the National Academy of Sciences,  **2017**, Vol. 114(39), pp. 10432-10437

Doiron, M., Gauthier, G. and Lévesque, E. *Trophic mismatch and its effects on the growth of young in an Arctic herbivore.* Global Change Biology,  **2015**, Vol. 21(12), pp. 4364-4376

Dornelas, M., Gotelli, N. J., McGill, B., Shimadzu, H., Moyes, F., Sievers, C. and Magurran, A. E. *Assemblage time series reveal biodiversity change but not systematic loss.* Science, **2014**, Vol. 344(6181), pp. 296-299.

Durant, J. M., Hjermann, D. Ø., Anker-Nilssen, T., Beaugrand, G., Mysterud, A., Pettorelli, N. and Stenseth, N. C. *Timing and abundance as key mechanisms affecting trophic interactions in variable environments*. Ecology Letters,  **2005**, Vol. 8(9), pp. 952-958.

Durant, J. M., Hjermann, D. Ø., Ottersen, G. and Stenseth, N. C. *Climate and the match or mismatch between predator requirements and resource availability*. Climate research, **2007**, Vol. 33(3), pp. 271-283

Edmondson, W. *Sixty years of Lake Washington: a curriculum vitae*. Lake and Reservoir Management, **1994**, Vol. 10(2), pp. 75-84.

Francis, T. B., Wolkovich, E. M., Scheuerell, M. D., Katz, S. L., Holmes, E. E. and Hampton, S. E. *Shifting regimes and changing interactions in the Lake Washington, USA, plankton community from 1962—1994*. PloS one, **2014**, Vol. 9(10), pp. e110363

Fritts, T. H. and Rodda, G. H. *The role of introduced species in the degradation of island ecosystems: a case history of Guam*. Annual review of Ecology and Systematics, **1998**, Vol. 29(1), pp. 113-140.

George, D.*The effect of nutrient enrichment and changes in the weather on the abundance of Daphnia in Esthwaite Water, Cumbria*. Freshwater Biology,  **2012**, Vol. 57(2), pp. 360-372

Gruner, D. S., Smith, J. E., Seabloom, E. W., Sandin, S. A., Ngai, J. T., Hillebrand, H., Harpole, W. S., Elser, J. J., Cleland, E. E., Bracken, M. E. and others. *A cross-system synthesis of consumer and nutrient resource control on producer biomass*. Ecology letters,  **2008**, Vol. 11(7), pp. 740-755

Gullett, P., Hatchwell, B. J., Robinson, R. A. and Evans, K. L. *Phenological indices of avian reproduction: cryptic shifts and prediction across large spatial and temporal scales*. Ecology and evolution,  **2013**, Vol. 3(7), pp. 1864-1877.

Hampton, S. E., Scheuerell, M. D. and Schindler, D. E. *Coalescence in the Lake Washington story: interaction strengths in a planktonic food web.* Limnology and Oceanography,  **2006**, Vol. 51(5), pp. 2042-2051.

Henrich-Gebhardt, S. G. *Temporal and spatial variation in food availability and its effects on fledgling size in the great tit*. Population biology of passerine birds **1990**, pp. 175-185.

Holliday, N. *Population ecology of winter moth (Operophtera brumata) on apple in relation to larval dispersal and time of bud burst*. Journal of Applied Ecology, JSTOR, **1977**, pp. 803-813.

Johansson, J. and Jonzén, N.*Game theory sheds new light on ecological responses to current climate change when phenology is historically mismatched*. Ecology letters,  **2012**, Vol. 15(8), pp. 881-888.

Johansson, J., Kristensen, N. P., Nilsson, J.-Å. and Jonzén, N.*The eco-evolutionary consequences of interspecific phenological asynchrony--a theoretical perspective*. Oikos,  **2015**, Vol. 124(1), pp. 102-112.

KELLERI, L. F. and VAN NOORDWIJK, A. J. *Effects of local environmental conditions*. Ardea, **1994**, Vol. 82, pp. 349-362

Kerby, J., Wilmers, C. and Post, E. *Climate change, phenology, and the nature of consumer--resource interactions: advancing the match/mismatch hypothesis*. Ohgushi, T., Schmitz, O. J. & Holt, R. D. *(ed.)* Trait-mediated indirect interactions: ecological and evolutionary perspectives, Cambridge University Press Cambridge, UK, **2012**, pp. 508-525.

Kharouba, H. M., Ehrlén, J., Gelman, A., Bolmgren, K., Allen, J. M., Travers, S. E. and Wolkovich, E. M. *Global shifts in the phenological synchrony of species interactions over recent decades*. Proceedings of the National Academy of Sciences,  **2018**, Vol. 115(20), pp. 5211-5216.

Lavergne, S., Mouquet, N., Thuiller, W. and Ronce, O. *Biodiversity and climate change: integrating evolutionary and ecological responses of species and communities*. Annual review of ecology, evolution, and systematics, Annual Reviews, **2010**, Vol. 41, pp. 321-350.

Law, T., Zhang, W., Zhao, J. and Arhonditsis, G. B. *Structural changes in lake functioning induced from nutrient loading and climate variability*. Ecological Modelling, **2009**, Vol. 220(7), pp. 979-997.

Leggett, W. and Deblois, E. *Recruitment in marine fishes: is it regulated by starvation and predation in the egg and larval stages?* Netherlands Journal of Sea Research, **1994**, Vol. 32(2), pp. 119-134.

Menzel, A., Sparks, T. H., Estrella, N., Koch, E., Aasa, A., Ahas, R., Alm-Kübler, K., Bissolli, P., Braslavská, O., Briede, A. and others *European phenological response to climate change matches the warming pattern*. Global change biology,  **2006**, Vol. 12(10), pp. 1969-1976.

Miller-Rushing, A. J., Høye, T. T., Inouye, D. W. and Post, E. *The effects of phenological mismatches on demography*. Philosophical Transactions of the Royal Society B: Biological Sciences,  **2010**, Vol. 365(1555), pp. 3177-3186.

Mouquet, N., Lagadeuc, Y., Devictor, V., Doyen, L., Duputié, A., Eveillard, D., Faure, D., Garnier, E., Gimenez, O., Huneman, P. and others. *Predictive ecology in a changing world*. Journal of Applied Ecology,  **2015**, Vol. 52(5), pp. 1293-1310.

Nakazawa, T. and Doi, H. *A perspective on match/mismatch of phenology in community contexts*. Oikos,  **2012**, Vol. 121(4), pp. 489-495.

O'Connor, M. I., Selig, E. R., Pinsky, M. L. and Altermatt, F.*Toward a conceptual synthesis for climate change responses*. Global Ecology and Biogeography,  **2012**, Vol. 21(7), pp. 693-703

Ovaskainen, O., Skorokhodova, S., Yakovleva, M., Sukhov, A., Kutenkov, A., Kutenkova, N., Shcherbakov, A., Meyke, E. and del Mar Delgado, M. *Community-level phenological response to climate change*. Proceedings of the National Academy of Sciences,  **2013**, Vol. 110(33), pp. 13434-13439

Philippart, C. J., van Aken, H. M., Beukema, J. J., Bos, O. G., Cadée, G. C. and Dekker, R.*Climate-related changes in recruitment of the bivalve Macoma balthica.* Limnology and Oceanography,  **2003**, Vol. 48(6), pp. 2171-2185

Plard, F., Gaillard, J.-M., Coulson, T., Hewison, A. M., Delorme, D., Warnant, C. and Bonenfant, C. *Mismatch between birth date and vegetation phenology slows the demography of roe deer*. PLoS biology, **2014**, Vol. 12(4), pp. e1001828

Post, E. and Forchhammer, M. C. *Climate change reduces reproductive success of an Arctic herbivore through trophic mismatch*. Philosophical Transactions of the Royal Society B: Biological Sciences, The Royal Society London, **2007**, Vol. 363(1501), pp. 2367-2373

Reed, T. E., Jenouvrier, S. and Visser, M. E. *Phenological mismatch strongly affects individual fitness but not population demography in a woodland passerine*. Journal of Animal Ecology,  **2013**, Vol. 82(1), pp. 131-144

Renner, S. S. and Zohner, C. M. *Climate change and phenological mismatch in trophic interactions among plants, insects, and vertebrates*. Annual Review of Ecology, Evolution, and Systematics, Annual Reviews, **2018**, Vol. 49, pp. 165-182.

Revilla, T. A., Encinas-Viso, F. and Loreau, M. *(A bit) Earlier or later is always better: Phenological shifts in consumer--resource interactions*. Theoretical ecology,  **2014**, Vol. 7(2), pp. 149-162.

Ricciardi, A., Neves, R. J. and Rasmussen, J. B. *Impending extinctions of North American freshwater mussels (Unionoida) following the zebra mussel (Dreissena polymorpha) invasion*. Journal of animal ecology,  **1998**, Vol. 67(4), pp. 613-619.

Rudolf, V. H. and Singh, M. *Disentangling climate change effects on species interactions: effects of temperature, phenological shifts, and body size*. Oecologia,  **2013**, Vol. 173(3), pp. 1043-1052

Sala, O. E., Chapin, F. S., Armesto, J. J., Berlow, E., Bloomfield, J., Dirzo, R., Huber-Sanwald, E., Huenneke, L. F., Jackson, R. B., Kinzig, A. and others. *Global biodiversity scenarios for the year 2100***.** Science, **2000**, Vol. 287(5459), pp. 1770-1774**.**

Samplonius, J. M., Kappers, E. F., Brands, S. and Both, C.*Phenological mismatch and ontogenetic diet shifts interactively affect offspring condition in a passerine.* Journal of Animal Ecology,  **2016**, Vol. 85(5), pp. 1255-1264.

Senner, N. R., Stager, M. and Sandercock, B. K. *Ecological mismatches are moderated by local conditions for two populations of a long-distance migratory bird.* Oikos,  **2017**, Vol. 126(1), pp. 61-72

Sgardeli, V., Zografou, K. and Halley, J. M. *Climate change versus ecological drift: assessing 13 years of turnover in a butterfly community*. Basic and applied ecology, **2016**, Vol. 17(4), pp. 283-290

Shurin, J. B., Borer, E. T., Seabloom, E. W., Anderson, K., Blanchette, C. A., Broitman, B., Cooper, S. D. and Halpern, B. S. *A cross-ecosystem comparison of the strength of trophic cascades*. Ecology letters,  **2002**, Vol. 5(6), pp. 785-791

Shurin, J. B., Gruner, D. S. and Hillebrand, H. *All wet or dried up? Real differences between aquatic and terrestrial food webs*. Proceedings of the Royal Society B: Biological Sciences, The Royal Society London, **2005**, Vol. 273(1582), pp. 1-9

Shurin, J. B. and Seabloom, E. W.*The strength of trophic cascades across ecosystems: predictions from allometry and energetics*. Journal of animal Ecology,  **2005**, Vol. 74(6), pp. 1029-1038

Singer, M. C. and Parmesan, C. *Phenological asynchrony between herbivorous insects and their hosts: signal of climate change or pre-existing adaptive strategy?* Philosophical Transactions of the Royal Society B: Biological Sciences,  **2010**, Vol. 365(1555), pp. 3161-3176

Solomon, S. *IPCC (2007): Climate change the physical science basis* **2007**

Thackeray, S. J.*Mismatch revisited: what is trophic mismatching from the perspective of the plankton?*Journal of Plankton Research, Oxford University Press, **2012**, Vol. 34(12), pp. 1001-1010

Thackeray, S. J., Henrys, P. A., Hemming, D., Bell, J. R., Botham, M. S., Burthe, S., Helaouet, P., Johns, D. G., Jones, I. D., Leech, D. I. and others. *Phenological sensitivity to climate across taxa and trophic levels*. Nature, **2016**, Vol. 535(7611), pp. 241

Tikkanen, O.-P. and Julkunen-Tiitto, R. *Phenological variation as protection against defoliating insects: the case of Quercus robur and Operophtera brumata*. Oecologia,  **2003**, Vol. 136(2), pp. 244-251

Tikkanen, O.-P., Niemelä, P. and Keränen, J. *Growth and development of a generalist insect herbivore, Operophtera brumata, on original and alternative host plants*. Oecologia,  **2000**, Vol. 122(4), pp. 529-536

Vatka, E., Orell, M. and RytkÖnen, S. *Warming climate advances breeding and improves synchrony of food demand and food availability in a boreal passerine*. Global Change Biology,  **2011**, Vol. 17(9), pp. 3002-3009

Vatka, E., Rytkönen, S. and Orell, M.*Does the temporal mismatch hypothesis match in boreal populations?*Oecologia,  **2014**, Vol. 176(2), pp. 595-605

Verschuren, D., Johnson, T. C., Kling, H. J., Edgington, D. N., Leavitt, P. R., Brown, E. T., Talbot, M. R. and Hecky, R. E. *History and timing of human impact on Lake Victoria, East Africa*. Proceedings of the Royal Society of London. Series B: Biological Sciences,  **2002**, Vol. 269(1488), pp. 289-294

Visser, M. E., Holleman, L. J. M. and Gienapp, P. *Shifts in caterpillar biomass phenology due to climate change and its impact on the breeding biology of an insectivorous bird*. Oecologia,  **2006**, Vol. 147(1), pp. 164-172.

Visser, M.E., Van Noordwijk, A.J., Tinbergen, J.M., and Lessells, C.M. *Warmer springs lead to mistimed reproduction in great tits (Parus major).*Philos. Trans. R. Soc. London B, **1998***,* 265:1867-1870.

Wiltshire, K. H., Malzahn, A. M., Wirtz, K., Greve, W., Janisch, S., Mangelsdorf, P., Manly, B. F. and Boersma, M. *Resilience of North Sea phytoplankton spring bloom dynamics: An analysis of long-term data at Helgoland Roads*. Limnology and Oceanography,  **2008**, Vol. 53(4), pp. 1294-1302

Winder, M. and Schindler, D. E.*Climate change uncouples trophic interactions in an aquatic ecosystem.*

Ecology,  **2004**, Vol. 85(8), pp. 2100-2106

Wolkovich, E., Cook, B., McLauchlan, K. and Davies, T. *Temporal ecology in the Anthropocene*. Ecology letters,  **2014**, Vol. 17(11), pp. 1365-1379

**Box 1. Glossary**

**Phenological synchrony-** changes in the relative timing of key life history events among pair-wise interacting species. Importantly, this term does not incorporate fitness consequences for either species.

**Phenological mismatch**- decreases in consumer fitness associated with changes in the relative timing of key life history activities among interacting species. This mismatch occurs between interacting species; it does not occur with the abiotic environment or due intraspecific processes.

**Cushing match-mismatch hypothesis**- this hypothesis postulates that there should be selective pressure for the consumer to temporally ‘match’ the peak of its energetic phase (i.e. reaching maximum fitness) with the peak of resource availability and if there is any change to the relative timing of the interaction, there will be a decrease in fitness for the consumer (i.e., a mismatch). At the curve’s limits, consumer fitness should fall to zero when the change in relative timing is sufficiently large. (Figure 1)

**Baseline-** a representative or benchmark time series of conditions that describes the species interaction. A baseline incorporates the historical range of variability in the relative timing of the interaction.

**Synchrony hypothesis-** a hypothesis that before climate change, the most energetically demanding phase of the consumer was at the same time of peak resource availability, and thus consumer fitness was at its maximum (i.e. a match).

**Asynchrony baseline** - a hypothesis that before climate change, the most energetically demanding phase of the consumer was not timed to the peak resource availability, and thus consumer fitness was not at its maximum. This could occur for various reasons (e.g., co-evolutionary arms race, other transient dynamics).

**Adaptive asynchrony** **hypothesis-** a hypothesis put forward by Singer and Parmesan (2010) that postulates that, in some systems, life-history trade-offs will promote asynchrony for many or most individuals in a population.

**Stationarity-** any stochastic process with a constant underlying probability distribution

**Life-history theory-** the study of the causes and consequences of the diversity of life history strategies based on evolutionary theory.

**Food web theory-** the study of the patterns, causes and consequences of food-web structure.

**Tables**

Table 1. A comparison across studies of the type of performance data collected for consumer and resource. We define a life-history study as one that collected data at the individual level and a food-web study as one that collected data at the population or community (i.e., across species) level*.* Counts in the table are numbers of individual pair-wise interactions (n=46).

|  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- |
|  | | | Resource performance | | | |  |
|  | Life-history | Food-web | |  |
| None | Individual | Population | Community | *Totals* |
| Consumer  performance | Life-history | Individual | 6 | 0 | 4 | 18 | *28* |
| Food-web | Population | 1 | 1 | 7 | 7 | *16* |
| Community | 0 | 0 | 1 | 1 | *2* |
|  | | **Total** | **7** | **1** | **12** | **26** | **46** |

Table 2. A comparison across studies of the type of performance data collected for the consumer across systems and taxonomic group. Counts in the table are numbers of individual pair-wise interactions (n=46).

|  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- |
|  | | System | | Taxonomic group | | | |
| Invertebrate | Vertebrate | | |
| Aquatic | Terrestrial | Fish | Bird | Mammal |
| Level of consumer  performance | Individual | 5 | 23 | 1 | 2 | 19 | 6 |
| Population | 14 | 2 | 9 | 4 | 2 | 1 |
| Community | 1 | 1 | 2 | 0 | 0 | 0 |
| **Sub-totals** | **20** | **26** | **12** | **6** | **21** | **7** |
|  | **Totals** |  | **46** |  |  |  | **46** |

**Figures**

Figure 1. Conceptualization of the Cushing match-mismatch hypothesis represented by the curve in panel (a). The hypothesis postulates that a consumer should temporally ‘match’ the peak of its energetic phase with the peak of resource availability and thus have the highest fitness (a,c) and if there is any change to the relative timing of the interaction (e.g., because the consumer shifts its phenology earlier (b) or later (d) relative to the resource), there will be a decrease in the consumer’s fitness (i.e., a mismatch) (a). At the curve’s limits, consumer fitness should fall to zero when the change in relative timing is sufficiently large (a). Curves in panels b-d represent the seasonal changes in the abundance of the consumer (black) and resource (red) where during some part of the year abundance declines to zero.

Figure 2. A simplified flow diagram for forecasting climate change effects on consumer fitness as predicted by the Cushing hypothesis. First, both major assumptions must be met; if, for example, consumer density is the major controller on its own fitness, then no further work is warranted. If both assumptions are met, forecasting requires both knowledge of the ultimate mechanisms (which often relate to testing the 1st assumption) and the pre-climate change baseline. In some cases, the mechanism predicts the pre-climate change baseline (i.e., in both our examples); we colour mechanisms in whether they predict both pre-climate change synchrony and asynchrony depending on specifics (purple), synchrony (blue) or asynchrony (red). Note that multiple mechanisms may operate in many systems, and that this figure is in no way exhaustive, but rather serves to highlight a pathway that researchers can take.

Figure 3. Conceptualization of key assumptions, and resulting implications for climate change predictions, often made about the Cushing hypothesis when pre-climate change baselines are not defined. (a) Differences in the phenological time-series of a consumer-resource interaction, where red represents the resource and black represents the consumer, during conditions of stationarity; when the environment becomes non-stationary (shown here just after 1980) the consumer and resource each can shift in varying directions, representing the range of recent documented phenological shifts (e.g. Thackeray et al. 2016), leading potentially to shifts in synchrony. (b) Most studies in the current literature assume that consumer fitness was highest before climate change (i.e., a match; synchrony hypothesis) leading to a ‘synchrony baseline’. However, an alternative hypothesis put forward by Singer and Parmesan (2010) (i.e., what they term as the ‘adaptive asynchrony’ hypothesis) postulates that conditions before climate change may not represent a ‘match’ in the system for many individuals. This hypothesis may lead to asynchrony as a pre-climate change baseline (see c) or a population where few individuals are matched; we show this latter possibility here (i.e. ‘adaptive asynchrony hypothesis with synchrony baseline’; our representation of this hypothesis is at the population level). The implications for climate change predictions for the two hypotheses are illustrated: If the synchrony baseline is supported, then climate change will necessarily lead to declines in consumer fitness. If the ‘adaptive asynchrony hypothesis’ with a synchrony baseline is supported, climate change may not lead to large declines in consumer fitness. (c) Without establishing a pre-climate change baseline and defining where an interaction falls along a curve, it is difficult to predict how climate change-driven changes to the relative timing of the interaction may affect consumer fitness. For example, with an asynchrony baseline, climate change could lead to an increase or decrease, or to varying magnitudes, in consumer fitness depending on how the relative timing of the interaction changes. For panels b and c, blue boxes represent the range of conditions detected in the system over a long time period.

Figure 4. Case study demonstrating the integration of experimental (a) and observational data (b) relating to the Cushing hypothesis in a single system- the winter moth (*Operophtera brumata*) and oak (*Quercus robur*). (a) The results from two related experiments (green, red points) where the authors manipulated the number of days that neonates (i.e. early instar larvae) spent without food (green points; first experiment) and the emergence times of larvae relative to budburst (red points; second experiment). Raw data was obtained from Tikkanen and Julkunen-Tiitto (2003; Figure 3). See Appendix for more details. (b) Inter-annual variation in relative timing between median egg hatch date of *O. brumata* and the median bud opening date of *Q. robur* from 1996-2005 in the Netherlands. Horizontal error bars represent the lower and upper quartiles of the data. Raw data from the observational study was retrieved from VanAsch and Visser 2007 Figure 2. For both panels, negative values along the x-axis denote where egg hatching occurred before bud opening (i.e. time without food), whereas positive values indicate egg hatching occurred after bud opening (i.e. time with food). While the experimental data from panel (a) shows support for the Cushing hypothesis, and the ultimate mechanism in this system suggests that the pre-climate change baseline was likely synchrony (Figure 2), empirical data from a 10 year post-climate change time-series (panel (b) only covers a small portion of the Cushing curve. This makes it difficult to define the pre-climate change baseline and thus accurately predict how the performance of *O. brumata* will be affected by changes in phenological synchrony due to climate change (Figure 2).

Figure 1.

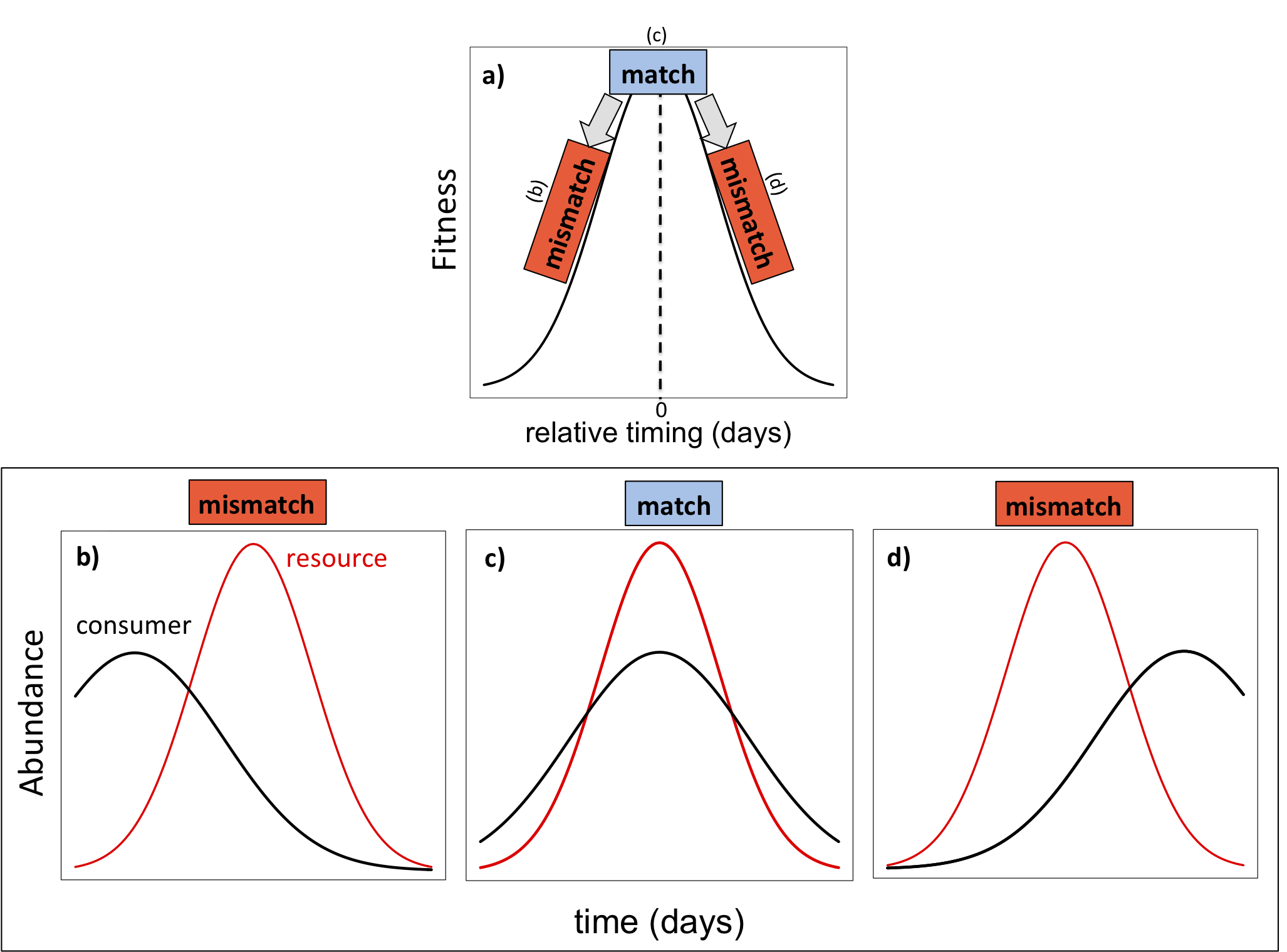


Figure 2.

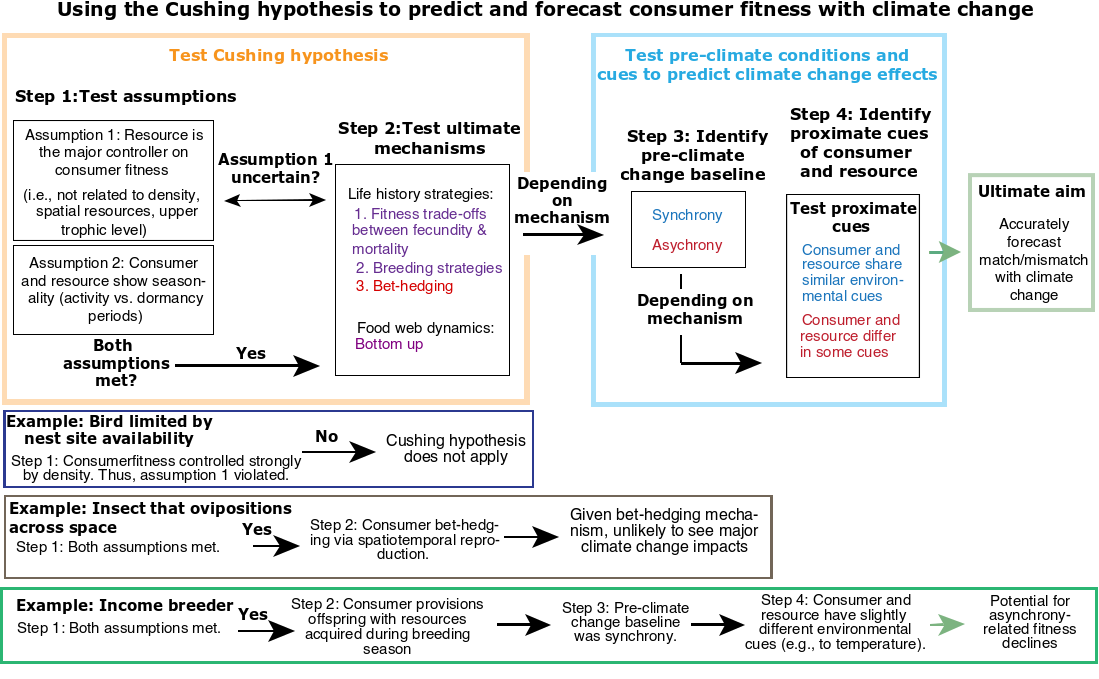


Figure 3.

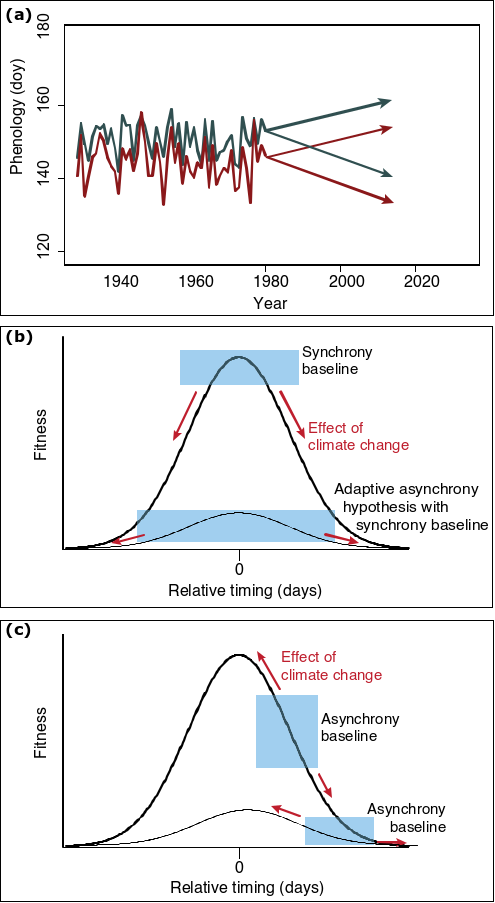
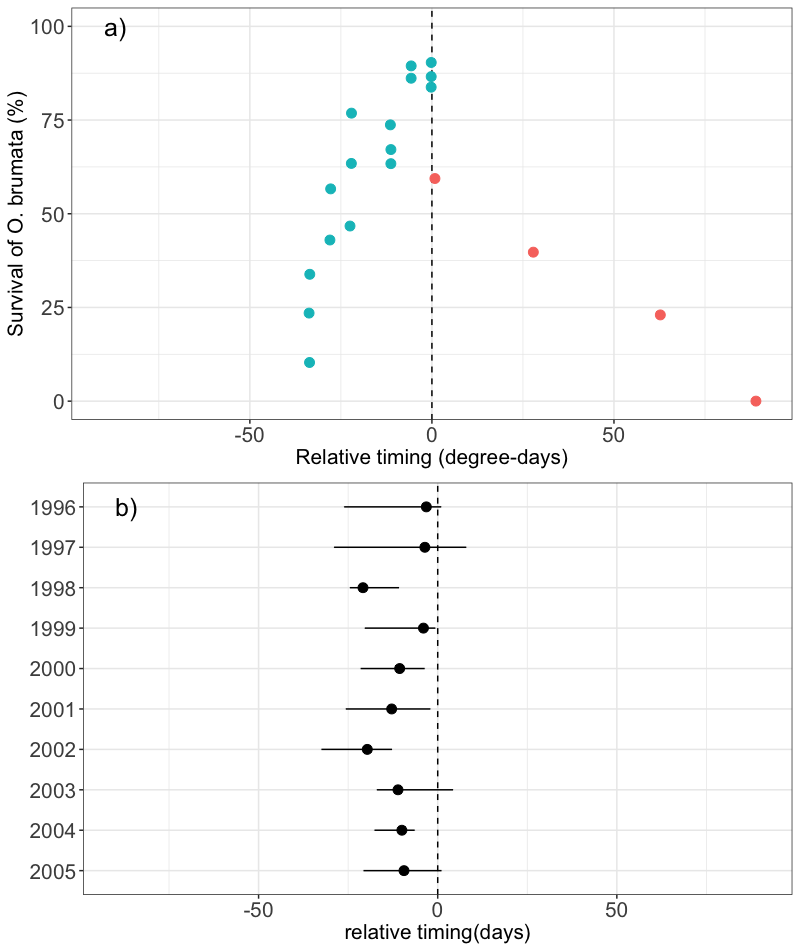


Figure 4.



**Appendix**

*Literature search*

We located papers relating phenological data from trophic interactions to fitness and/or performance of the consumer and/or the resource by conducting keyword searches in ISI Web of Science published up to June 2017. Keywords included phenolog\* AND mismatch\* OR synchron\* AND interact\* AND (fitness\* OR performance\*). Our initial search netted 2906 papers so we further refined our search by excluding categories that included engineering and computer science. This resulted in 393 papers. From these, we focused on observational studies and excluded studies if they did not: (1) measure phenology directly (e.g. used derived measures of phenology, such as NDVI or spring temperature); (2) measure phenology of at least one of the species; (3) quantitatively link consumer fitness or performance to the relative timing between consumer and resource; and (4) explicitly state that the two species interacted (e.g. specifying type of interaction). Estimates of consumer fitness or performance (i.e., growth or development) had to be direct estimates (e.g. we did not include studies that used diet proportions to measure consumer performance) and included a range of proxies, for example, growth, survival, abundance. To ensure a reasonable sample size and to include studies across different major biomes, we included interactions that were resolved to the family-level and below. To reduce redundancy within systems, we only kept studies that were unique across pairs-location-year combinations. If more than one measure of phenology was included, we chose the one used by the authors to calculate mismatch and examine its impact on performance.

Our final review included 43 studies with 46 pair-wise species interactions (3 studies had 2 interactions). These studies encompassed terrestrial, marine and freshwater ecosystems as well as a large latitudinal gradient.

*Summary of studies and interactions*

We classified studies as ‘climate change’ or ‘fundamental’ depending on whether they mentioned climate change in the abstract or in the introduction of the paper. Based on the type of data collected for the consumer and resource, we classified these studies as life history (i.e. one that collected data at the individual level) or food web-based (i.e. one that collected data at the population or community (i.e., across species)). To determine whether studies had the potential to define pre-climate change baselines, we measured the study’s time span and years of data based on the years where phenology data was available for both the consumer and resource, and consumer performance data was available.

*Additional details for Figure 4*

Raw data for panel (a) was obtained from Figure 3 in Tikkanen and Julkunen-Tiitto (2003). The data come from two related experiments where the authors manipulated the number of days that neonates (i.e. early instar larvae) spent without food (first experiment) and the emergence times of larvae relative to budburst (second experiment). In the first experiment, there were six groups of 30 larvae that spent 0, 5.5, 11, 22, 27,5 and 33 degree-days without food. In the second experiment, there were four cohorts, each separated by intervals of 3-5 days. All *O. brumata* eggs and larvae originated from laboratory stock originally from Turku, Finland whereas the foliage originated from trees near Banchory, NW Scotland.

*References for systematic literature review*

Adrian, R., Wilhelm, S. and Gerten, D. *Life-history traits of lake plankton species may govern their phenological response to climate warming*. Global change biology, **2006**, Vol. 12(4), pp. 652-661.

Arula, T., Gröger, J., Ojaveer, H. and Simm, M.*Shifts in the spring herring (Clupea harengus membras) larvae and related environment in the Eastern Baltic Sea over the past 50 years*. PloS one, Public Library of Science, **2014**, Vol. 9(3), pp. e91304

Atkinson, A., Harmer, R. A., Widdicombe, C. E., McEvoy, A. J., Smyth, T. J., Cummings, D. G., Somerfield, P. J., Maud, J. L. and McConville, K. *Questioning the role of phenology shifts and trophic mismatching in a planktonic food web*. Progress in Oceanography, **2015**, Vol. 137, pp. 498-512.

Blackett, M., Lucas, C. H., Harmer, R. A. and Licandro, P. *Population ecology of Muggiaea atlantica (Cnidaria, Siphonophora) in the Western English Channel.* Marine Ecology Progress Series, **2015**, Vol. 535, pp. 129-144

Borcherding, J., Beeck, P., DeAngelis, D. L. and Scharf, W. R. *Match or mismatch: the influence of phenology on size-dependent life history and divergence in population structure*. Journal of animal ecology,  **2010**, Vol. 79(5), pp. 1101-1112

Cresswell, W. and Mccleery, R. *How great tits maintain synchronization of their hatch date with food supply in response to long-term variability in temperature*. Journal of Animal Ecology,  **2003**, Vol. 72(2), pp. 356-366

Dekker, R. and Beukema, J. *Phenology of abundance of bivalve spat and of their epibenthic predators: limited evidence for mismatches after cold winters*. Marine Ecology Progress Series, **2014**, Vol. 513, pp. 17-27

Dessborn, L., Elmberg, J., Nummi, P., Pöysä, H. and Sjöberg, K. *Hatching in dabbling ducks and emergence in chironomids: a case of predator--prey synchrony?* Hydrobiologia,  **2009**, Vol. 636(1), pp. 319-329

Dias, P. C. and Blondel, J. *Breeding time, food supply and fitness components of Blue Tits Parus caeruleus in Mediterranean habitats*. Ibis,  **1996**, Vol. 138(4), pp. 644-649.

Dunn, P. O., Winkler, D. W., Whittingham, L. A., Hannon, S. J. and Robertson, R. J. *A test of the mismatch hypothesis: How is timing of reproduction related to food abundance in an aerial insectivore?* Ecology,  **2011**, Vol. 92(2), pp. 450-461

Fortierl, L. and Gilbert, M.*The match/mismatch hypothesis and the feeding success of fish larvae in ice-covered southeastern Hudson Bay* Mar. Ecol. Prog. Ser, **1995**, Vol. 120, pp. 11-27

George, D.*The effect of nutrient enrichment and changes in the weather on the abundance of Daphnia in Esthwaite Water, Cumbria.* Freshwater Biology,  **2012**, Vol. 57(2), pp. 360-372

Gullett, P., Hatchwell, B. J., Robinson, R. A. and Evans, K. L. *Phenological indices of avian reproduction: cryptic shifts and prediction across large spatial and temporal scales*. Ecology and evolution,  **2013**, Vol. 3(7), pp. 1864-1877.

Hipfner, J. M. *Matches and mismatches: ocean climate, prey phenology and breeding success in a zooplanktivorous seabird*. Marine Ecology Progress Series, **2008**, Vol. 368, pp. 295-304

Jolley, J. C., Willis, D. W. and Holland, R. S. *Match--mismatch regulation for bluegill and yellow perch larvae and their prey in Sandhill lakes*. Journal of Fish and Wildlife Management, **2010**, Vol. 1(2), pp. 73-85

Kerby, J. and Post, E. *Capital and income breeding traits differentiate trophic match--mismatch dynamics in large herbivores*. Philosophical Transactions of the Royal Society B: Biological Sciences,  **2013**, Vol. 368(1624), pp. 20120484.

Kerby, J. T. and Post, E. *Advancing plant phenology and reduced herbivore production in a terrestrial system associated with sea ice decline*. Nature Communications, **2013**, Vol. 4, pp. 2514

Kourkgy, C., Garel, M., Appolinaire, J., Loison, A. and To\igo, C.*Onset of autumn shapes the timing of birth in Pyrenean chamois more than onset of spring***.** Journal of Animal Ecology,  **2016**, Vol. 85(2), pp. 581-590

Lany, N. K., Ayres, M. P., Stange, E. E., Sillett, T. S., Rodenhouse, N. L. and Holmes, R. T.*Breeding timed to maximize reproductive success for a migratory songbird: The importance of phenological asynchrony.*Oikos,  **2016**, Vol. 125(5), pp. 656-666

McKinnon, L., Picotin, M., Bolduc, E., Juillet, C. and Bêty, J. *Timing of breeding, peak food availability, and effects of mismatch on chick growth in birds nesting in the High Arctic*. Canadian Journal of Zoology, **2012**, Vol. 90(8), pp. 961-971

Mortensen, L. O., Schmidt, N. M., Høye, T. T., Damgaard, C. and Forchhammer, M. C. *Analysis of trophic interactions reveals highly plastic response to climate change in a tri-trophic high-Arctic ecosystem*. Polar Biology,  **2016**, Vol. 39(8), pp. 1467-1478

Pakanen, V.-M., Orell, M., Vatka, E., Rytkönen, S. and Broggi, J. *Different ultimate factors define timing of breeding in two related species*. PloS one, Public Library of Science, **2016**, Vol. 11(9), pp. e0162643

Philippart, C. J., van Aken, H. M., Beukema, J. J., Bos, O. G., Cadée, G. C. and Dekker, R. *Climate-related changes in recruitment of the bivalve Macoma balthica.* Limnology and Oceanography,  **2003**, Vol. 48(6), pp. 2171-2185.

Plard, F., Gaillard, J.-M., Coulson, T., Hewison, A. M., Delorme, D., Warnant, C. and Bonenfant, C. *Mismatch between birth date and vegetation phenology slows the demography of roe deer*. PLoS biology, Public Library of Science, **2014**, Vol. 12(4), pp. e1001828.

Post, E. and Forchhammer, M. C. *Climate change reduces reproductive success of an Arctic herbivore through trophic mismatch.* Philosophical Transactions of the Royal Society B: Biological Sciences, The Royal Society London, **2007**, Vol. 363(1501), pp. 2367-2373.

Reed, T. E., Grøtan, V., Jenouvrier, S., Sæther, B.-E. and Visser, M. E. *Population growth in a wild bird is buffered against phenological mismatch.* Science, **2013**, Vol. 340(6131), pp. 488-491.

Reed, T. E., Jenouvrier, S. and Visser, M. E. *Phenological mismatch strongly affects individual fitness but not population demography in a woodland passerine.* Journal of Animal Ecology,  **2013**, Vol. 82(1), pp. 131-144.

Régnier, T., Gibb, F. and Wright, P. *Importance of trophic mismatch in a winter-hatching species: evidence from lesser sandeel.* Marine Ecology Progress Series, **2017**, Vol. 567, pp. 185-197

Reneerkens, J., Schmidt, N. M., Gilg, O., Hansen, J., Hansen, L. H., Moreau, J. and Piersma, T. *Effects of food abundance and early clutch predation on reproductive timing in a high Arctic shorebird exposed to advancements in arthropod abundance*. Ecology and Evolution,  **2016**, Vol. 6(20), pp. 7375-7386

Salamolard, M., Butet, A., Leroux, A. and Bretagnolle, V.*Responses of an avian predator to variations in prey density at a temperate latitude.* Ecology,  **2000**, Vol. 81(9), pp. 2428-2441

Seebens, H., Einsle, U. and Straile, D. *Copepod life cycle adaptations and success in response to phytoplankton spring bloom phenology*. Global Change Biology,  **2009**, Vol. 15(6), pp. 1394-1404

Senner, N. R., Stager, M. and Sandercock, B. K.*Ecological mismatches are moderated by local conditions for two populations of a long-distance migratory bird*. Oikos,  **2017**, Vol. 126(1), pp. 61-72

Sergeant, C. J., Armstrong, J. B. and Ward, E. J. *Predator-prey migration phenologies remain synchronised in a warming catchment*. Freshwater Biology,  **2015**, Vol. 60(4), pp. 724-732

Shultz, M. T., Piatt, J. F., Harding, A. M., Kettle, A. B. and Van Pelt, T. I. *Timing of breeding and reproductive performance in murres and kittiwakes reflect mismatched seasonal prey dynamics*. Marine Ecology Progress Series, **2009**, Vol. 393, pp. 247-258

Sullivan, B. K., Costello, J. H. and Van Keuren, D. *Seasonality of the copepods Acartia hudsonica and Acartia tonsa in Narragansett Bay, RI, USA during a period of climate change*. Estuarine, Coastal and Shelf Science, **2007**, Vol. 73(1-2), pp. 259-267

Thomas, D. W., Blondel, J., Perret, P., Lambrechts, M. M. and Speakman, J. R.*Energetic and fitness costs of mismatching resource supply and demand in seasonally breeding birds.* Science, American Association for the Advancement of Science, **2001**, Vol. 291(5513), pp. 2598-2600

Vatka, E., Orell, M. and RytkÖnen, S. *Warming climate advances breeding and improves synchrony of food demand and food availability in a boreal passerine.* Global Change Biology,  **2011**, Vol. 17(9), pp. 3002-3009.

Vatka, E., Rytkönen, S. and Orell, M.*Does the temporal mismatch hypothesis match in boreal populations?*Oecologia,  **2014**, Vol. 176(2), pp. 595-605

Visser, M. E., Gienapp, P., Husby, A., Morrisey, M., de la Hera, I., Pulido, F. and Both, C.*Effects of spring temperatures on the strength of selection on timing of reproduction in a long-distance migratory bird***.** PLoS Biology, Public Library of Science, **2015**, Vol. 13(4), pp. e1002120

Watanuki, Y., Ito, M., Deguchi, T. and Minobe, S.*Climate-forced seasonal mismatch between the hatching of rhinoceros auklets and the availability of anchovy*. Marine Ecology Progress Series, **2009**, Vol. 393, pp. 259-271

Wiltshire, K. H., Malzahn, A. M., Wirtz, K., Greve, W., Janisch, S., Mangelsdorf, P., Manly, B. F. and Boersma, M. *Resilience of North Sea phytoplankton spring bloom dynamics: An analysis of long-term data at Helgoland Roads*. Limnology and Oceanography,  **2008**, Vol. 53(4), pp. 1294-1302

Winder, M. and Schindler, D. E.*Climate change uncouples trophic interactions in an aquatic ecosystem*. Ecology,  **2004**, Vol. 85(8), pp. 2100-2106